



# Effect of emission control measures on ozone concentrations in Hangzhou during G20 meeting in 2016

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## HIGHLIGHTS

- MDA8 O<sub>3</sub> concentrations in Hangzhou during G20 exceeded national air quality standard.
- Emission control measures during G20 reduced MDA8 O<sub>3</sub> level in Hangzhou by 11.7%.
- Emission controls in industry and transportation sectors are most effective for controlling O<sub>3</sub>.

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## ABSTRACT

The effect of emission control measures on ozone (O<sub>3</sub>) concentrations in Hangzhou during G20 (The Group of Twenty Finance Ministers and Central Bank Governors) meeting during 24 August to 6 September of 2016 was evaluated using the nested version of a global chemical transport model. During G20, observed concentrations of PM<sub>10</sub>, PM<sub>2.5</sub>, SO<sub>2</sub>, NO<sub>2</sub>, and CO were all below national air quality standards, whereas those of MDA8 O<sub>3</sub> were above national standard (with an averaged value of 160.2 μg m<sup>-3</sup>) but had a decreasing trend. Model sensitivity studies show that, MDA8 O<sub>3</sub> concentrations in Hangzhou during G20 were reduced by 11.3 μg m<sup>-3</sup> (6.8%), 14.8 μg m<sup>-3</sup> (8.9%), and 19.5 μg m<sup>-3</sup> (11.7%) with emission control measures in the core area, Zhejiang province, and the Yangtze River Delta (YRD) region, respectively, indicating that control measures were the most effective when carried out jointly in YRD. Considering the ratios of NO<sub>x</sub> to VOCs during G20, Hangzhou and most areas of Zhejiang province were in transitional regime; reductions in either NO<sub>x</sub> or VOCs could reduce O<sub>3</sub> concentrations. We also quantified how sensitive O<sub>3</sub> concentrations respond to emission reductions in sectors of industry, power, residential and transportation in the whole of YRD during G20. The removal of emissions in industry and transportation sectors would lead to the largest reductions of 17.6 μg m<sup>-3</sup> (10.5%) and 12.3 μg m<sup>-3</sup> (7.4%) in MDA8 O<sub>3</sub> concentrations in Hangzhou during G20, respectively. This study has important implications for the control of high O<sub>3</sub> levels in eastern China.

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## 1. Introduction

Air quality in China has drawn a lot of attention because of the high concentrations of PM<sub>2.5</sub> and ozone (O<sub>3</sub>) resulted from its rapid economic development. The Chinese government launched the 'Air Pollution Prevention and Control Action Plan' to reduce anthropogenic emissions to improve air quality in 2013. According to the 'Environmental and Ecological Status Bulletins in China' ([http://](http://www.mee.gov.cn/hjzl/)

[www.mee.gov.cn/hjzl/](http://www.mee.gov.cn/hjzl/)), from 2013 to 2017, mean concentrations of PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub> and CO in 74 major cities decreased by 34.7%, 32.2%, 57.5%, 9.1% and 32.0%, respectively, but the 90th percentile concentration of the maximum daily 8-h average of O<sub>3</sub> concentration (MDA8 O<sub>3</sub>) averaged over these cities increased by 20.1%. Under the background of increasingly serious O<sub>3</sub> pollution in China, G20 meeting offers a good opportunity to study whether O<sub>3</sub> concentrations can be controlled effectively in summer in the polluted eastern China.

In recent years, several significant events were held in China successfully, such as the Beijing Olympic Games in 2008, the Shanghai World Expo in 2010, the Asia-Pacific Economic

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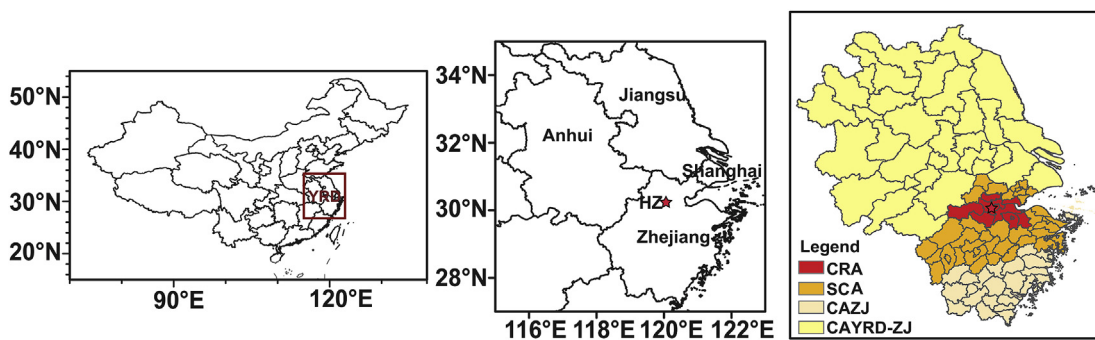
Cooperation (APEC) conference in 2014, the Victory Day Parade (V-Day Parade) in 2015, and the Group of Twenty Finance Ministers and Central Bank Governors (G20) summit in 2016. For purpose of ensuring good air quality during these important events, air quality control measures were implemented before and during the events, which were proved to be very effective by measured concentrations of air pollutants. A number of studies were carried out to quantify the effectiveness of the control measures during these events. For example, Gao et al. (2011) reported by using the Weather Research and Forecasting model coupled to Chemistry (WRF-Chem) that PM<sub>2.5</sub> concentrations in Beijing decreased by about 30%, as a result of the reductions in anthropogenic emissions by about 50%, 35%, and 10% in Beijing, Hebei province, and other nearby regions, respectively, during the Beijing Olympic Games in 8–24 August 2008. Wang et al. (2010) used the fifth-generation NCAR/Penn State Mesoscale Model and Community Multi-scale Air Quality (MM5-CMAQ) model to simulate the reductions of pollutants during the Beijing Olympic Games and showed that, with emissions of SO<sub>2</sub>, NO<sub>x</sub>, PM<sub>10</sub>, and NMVOCs reduced by 41%, 47%, 55%, and 57%, respectively, PM<sub>2.5</sub> concentrations during the Olympic Games were reduced by about 60% compared to concentrations in June of 2008. Huang et al. (2017) quantified the effects of local emission reductions during the 2014 Nanjing Youth Olympic Games (NYOG) by using the Weather Research and Forecast and Community Multi-scale Air Quality (WRF-CMAQ) model. Under an unfavorable weather condition during NYOG (1–31 August, 2014), simulated mean concentrations of SO<sub>2</sub>, NO<sub>2</sub>, PM<sub>10</sub>, PM<sub>2.5</sub>, and CO decreased by 24.6%, 12.1%, 15.1%, 8.1%, and 7.2% when emissions of these species were reduced by 25.0%, 15.0%, 42.8%, 32.6%, and 20.0%, respectively. Liang et al. (2017) estimated the effectiveness of emission reductions during 2014 APEC meeting (3–12 November, 2014) and 2015 Victory Day Parade (20 August to 3 September, 2015) by using a generalized linear regression model (GLM). They found that the meteorological conditions and pollution control strategies contributed, respectively, 30% and 28% to the reductions of PM<sub>2.5</sub> concentrations during APEC and 38% and 25% to those during the Victory Parade.

Although the above-mentioned emission reductions were very effective in reducing PM<sub>2.5</sub> concentrations during those events, previous studies revealed that O<sub>3</sub> concentrations increased or remained unchanged during some events. For instance, Tang et al. (2017) examined the sensitivity of O<sub>3</sub> to emission reductions during the Beijing Olympic Games by using the MM5-CMAQ model. They found that the average O<sub>3</sub> concentration increased by 8 ppbv in urban Beijing where NO<sub>x</sub> decreased by about 15 ppbv during the Olympics, indicating the impact of titration effect. Xu et al. (2016) also reported on the basis of observations that emission

reductions during Beijing Olympics had a small impact on O<sub>3</sub> concentrations; averaged concentration of O<sub>3</sub> in Beijing and its surroundings was reduced by about 1% with a 33% decrease in NO<sub>x</sub> concentration, suggesting that O<sub>3</sub> production was VOC-limited. Guo et al. (2016) used the WRF-Chem model to evaluate the effect of district-joint emission controls on air quality during 2014 APEC meeting. Emission reductions from sectors of industry, residential, and transportation in urban Beijing and Huairou were 50%, 40% and 40%, respectively, which led to decreases in concentrations of PM<sub>2.5</sub>, PM<sub>10</sub>, NO<sub>2</sub>, and CO by 22%, 24%, 10% and 22%, respectively, whereas mean O<sub>3</sub> concentration increased by 24% due to the reductions in NO<sub>2</sub>. Li et al. (2017c) reported similar results; the emission control during APEC increased O<sub>3</sub> concentration by 58.2% because the degree of NO<sub>x</sub> reduction exceeded that of VOCs.

The Group of twenty (G20) was formally established in 1999, which comprises 19 countries plus the European Union. The G20 summit provided the opportunity for G20 members to explore approaches to the economic cooperation for the world economy. G20 took place in Hangzhou (119.8°E, 30.4°N), the capital city of Zhejiang province in the Yangtze River Delta region (YRD), during 4–5 September, 2016. In order to ensure the ‘blue sky’ during the conference, the government of Zhejiang province carried out a series of district-joint air quality control measures which took Hangzhou as the core city and radiated to other cities in Zhejiang and neighbor provinces. The areas that implemented the air pollution control measures included the core area (CRA), the strictly controlled area (SCA), the controlled area in Zhejiang (CAZJ) and the controlled area in Yangtze River Delta except Zhejiang (CAYRD-ZJ) (Fig. 1). Most stringent control measures were carried out in CRA, followed by the other three areas. The measures aimed to reduce emissions from industry, transportation, residential and power sectors (Yu et al., 2018). With the emission control measures, concentrations of all routine pollutants except for O<sub>3</sub> (PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub>, and CO) during G20 met the National air quality standards.

There were several studies about the effectiveness of emission control measures during G20. Su et al. (2017) used lidar data to analyze the difference in concentrations of aerosols and O<sub>3</sub> between G20 and post-G20 period, and found that aerosol extinction coefficient during G20 was 50% lower than that in post-G20 period in the lower lidar layer, while O<sub>3</sub> concentration during G20 was 37% higher than that in post-G20 period. Mao and Hu (2017) reported that, compared with September 2015 and August 2016, the monthly average Air Quality Index (AQI) during G20 were lower by 35 and 25 in the core area and lower by 20 and 25 in the strictly controlled area. Li et al. (2017b) used WRF-CMAQ model to evaluate the effect of emission controls on PM<sub>2.5</sub> and O<sub>3</sub> and concluded that



**Fig. 1.** The areas implemented control measures during G20 meeting. YRD consists of Zhejiang province, Jiangsu province, Anhui province and Shanghai (left and middle panels). The core area (CRA), the strictly controlled area (SCA), the controlled area in Zhejiang (CAZJ) and the controlled area in Yangtze River Delta except Zhejiang (CAYRD-ZJ) are shown in the right panel. The red star in the right two panels indicates the location of Hangzhou. (For interpretation of the references to color in this figure legend, the reader is referred to the Web version of this article.)

simulated concentrations of PM<sub>2.5</sub> and O<sub>3</sub> decreased by 56% and 25%, respectively, as a result of reductions in emissions from power plants and industry by 40% in Shanghai, Jiangsu and Anhui provinces and by 50% in Zhejiang province during August 24 to September 6. However, previous studies about G20 did not examine the effectiveness of local and joint emission control measures for O<sub>3</sub> as well as the sensitivity of O<sub>3</sub> concentrations to emission reductions in different sectors. We present results of such a study here by using a global 3-D model of atmospheric chemistry driven by meteorological input from the Goddard Earth Observing System (GEOS-Chem).

The descriptions of the GEOS-Chem model, emissions, observations, and numerical experiments are presented in Section 2. Section 3 presents the analyses of observed meteorological conditions and concentrations of pollutants, model evaluation, as well as the simulated changes in O<sub>3</sub> with reductions in emissions. Section 4 summarizes the main conclusions.

## 2. Method

### 2.1. The GEOS-Chem model

The simulation of air quality during G20 is carried out by using the GEOS-Chem model version11-01 ([http://wiki.seas.harvard.edu/geos-chem/index.php/GEOS-Chem\\_v11-01](http://wiki.seas.harvard.edu/geos-chem/index.php/GEOS-Chem_v11-01)), which is a global 3-D chemical transport model (CTM) driven by meteorological fields from the Modern-Era Retrospective Analysis for Research and Applications, version 2 (MERRA-2) provided by the Global Modelling and Assimilation Office (GMAO). The model includes fully coupled O<sub>3</sub>-NO<sub>x</sub>-hydrocarbon chemistry (Bey et al., 2001; Park et al., 2004) and aerosols including sulfate (Park et al., 2004), nitrate (Pye et al., 2009), ammonium, black carbon and organic carbon (Park et al., 2003), mineral dust (Fairlie et al., 2007), and sea salt (Alexander et al., 2005). Photolysis rates are calculated by Fast-JX scheme (Bian and Prather, 2002). Heterogeneous reactions of aerosols, such as the irreversible absorption of NO<sub>3</sub> and NO<sub>2</sub> on wet aerosols (Jacob, 2000), hydrolysis of N<sub>2</sub>O<sub>5</sub> (Evans and Jacob, 2005), and the uptake of HO<sub>2</sub> by aerosols (Thornton et al., 2008) are included. Wet deposition in the GEOS-Chem model, including scavenging in convective updrafts, rainout, and wash out, follows the scheme described by Liu et al. (2001) and applies only to soluble aerosols

and gases. Dry deposition is computed based on the resistance-in-series scheme of Wesely (1989). The nested domain for Asia (70°–150°E, 10°S–55°N) has a horizontal resolution of 0.5° latitude by 0.625° longitude and 47 vertical layers up to 0.01 hPa. Tracer concentrations at the lateral boundaries are provided by a global GEOS-Chem simulation at 2° latitude by 2.5° longitude horizontal resolution. The GEOS-Chem model has been used extensively for studying air quality in China, including haze pollution (Huang et al., 2014; Zhang et al., 2015, 2016; Yang et al., 2016; Li et al., 2016a; Qiu et al., 2017; Cai et al., 2017), O<sub>3</sub> air quality (Fu et al., 2009; Yang et al., 2014; Zhu and Liao, 2016; Ni et al., 2018a), and the effectiveness of emission controls during the Olympics (Wang et al., 2009) and APEC meeting (Gu and Liao, 2016).

### 2.2. Emissions

Muti-resolution Emission Inventory for China (MEIC) for year 2016 is used in our simulations. The inventory has a resolution of 0.25° × 0.25° and is downloaded from [www.meicmodel.org/dataset-meic.html](http://www.meicmodel.org/dataset-meic.html). Fig. 2 shows the spatial distributions of NO<sub>x</sub> and anthropogenic non-methane volatile organic carbons (NMVOCs) for August of 2016. In YRD, regions with high emissions are Shanghai, southern Jiangsu, central Anhui and northeastern Zhejiang.

Table 1 shows anthropogenic emissions of O<sub>3</sub> precursors, aerosols and aerosol precursors in regions of CRA, SCA, CAZJ and CAYRD-ZJ during 24 August to 6 September of 2016 from the MEIC inventory. Emissions of CO, NO<sub>x</sub>, and NMVOCs were the highest among all species, which were 622.5 × 10<sup>3</sup> kg, 136.4 × 10<sup>3</sup> kg, and 78.9 × 10<sup>3</sup> kgC in YRD (YRD = CRA + SCA + CAZJ + CAYRD-ZJ), respectively. Emissions of NO<sub>x</sub> were mainly from industry (52.3 × 10<sup>3</sup> kg) and transportation (50.8 × 10<sup>3</sup> kg) sectors, accounting for 38.4% and 37.2% of total NO<sub>x</sub> emissions in YRD during G20, respectively. Sectors of industry, transportation, residential, and power contributed 68.3%, 0.2%, 10.5% and 21.0%, respectively, to total NMVOCs emissions in YRD during G20.

The percentages of emission reductions for different species and in different sectors for periods of 24–27 August and 28 August–6 September are taken from Environmental Quality Guarantee Scheme during G20 in 2016 (Yu et al., 2018) (Table 1). The industry sector had the largest percentage reductions in emissions during

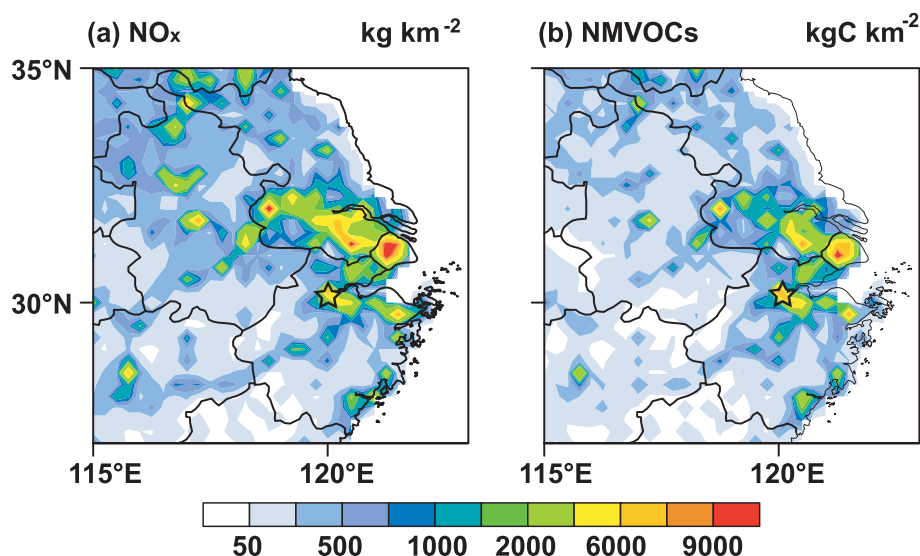


Fig. 2. Anthropogenic emissions of (a) NO<sub>x</sub> and (b) NMVOCs in YRD in August, 2016. The black star indicates the location of Hangzhou.

**Table 1**  
Total anthropogenic emissions of O<sub>3</sub> precursors, aerosols and aerosol precursors in the regions of CRA, Zhejiang province, and YRD for emission-controlled period of 24 August to 6 September (units: 10<sup>3</sup> kgC for NMVOCs and 10<sup>3</sup> kg for other species). Emissions are taken from MEIC for 2016, and emission reductions in percentages are taken from Environmental Quality Guarantee Scheme during G20 in 2016 (Yu et al., 2018).

Period		24–27 August				28 August - 6 September				
Sector		Industry	Power	Residential	Transport	Industry	Power	Residential	Transport	Total
Emissions in CRA	NO <sub>x</sub>	1.37	0.54	0.02	0.98	3.53	1.15	0.04	2.50	10.13
	NMVOCs	2.43	0.003	0.17	0.66	5.03	0.01	0.35	1.28	9.93
	CO	5.33	0.42	0.58	5.64	13.65	0.90	1.45	14.36	42.34
	SO <sub>2</sub>	0.76	0.16	0.02	0.05	1.94	0.35	0.06	0.13	3.47
	BC	0.04	0	0.01	0.03	0.1	0	0.02	0.08	0.28
	OC	0.03	0	0.02	0.01	0.08	0	0.04	0.03	0.21
	Reductions	100%	50%	50%	0%	100%	50%	50%	50%	–
Emissions in SCA	NO <sub>x</sub>	1.64	1.35	0.03	1.46	4.24	2.89	0.07	3.72	15.40
	NMVOCs	2.83	0.01	0.24	0.91	5.86	0.01	0.49	1.78	12.13
	CO	6.80	1.63	1.36	7.86	17.42	3.48	3.41	20.00	61.95
	SO <sub>2</sub>	0.91	0.4	0.04	0.08	2.33	0.85	0.09	0.19	4.89
	BC	0.05	0	0.02	0.05	0.12	0	0.04	0.12	0.40
	OC	0.04	0	0.05	0.02	0.1	0	0.12	0.05	0.38
	Reductions	50%	30%	30%	0%	50%	30%	30%	50%	–
Emissions in CAZJ	NO <sub>x</sub>	0.87	0.5	0.02	0.9	2.26	1.09	0.06	2.3	8.00
	NMVOCs	1.56	0.003	0.19	0.58	3.23	0.01	0.38	1.12	7.07
	CO	3.34	1.08	1.05	4.96	8.54	2.36	2.62	12.62	36.58
	SO <sub>2</sub>	0.48	0.13	0.03	0.05	1.23	0.28	0.07	0.12	2.39
	BC	0.02	0	0.01	0.03	0.06	0	0.03	0.07	0.22
	OC	0.02	0	0.04	0.01	0.05	0	0.09	0.03	0.24
	Reductions	50%	30%	30%	0%	50%	30%	30%	0%	–
Emissions in CAYRD-ZJ	NO <sub>x</sub>	10.58	7.5	0.41	10.96	27.81	16.61	1.03	27.94	102.84
	NMVOCs	10.63	0.04	2.14	3.38	22.29	0.08	4.3	6.86	49.72
	CO	65.17	8.94	31.04	30.98	169.12	19.82	77.59	78.98	481.63
	SO <sub>2</sub>	5.46	2.13	0.3	0.44	14.18	4.7	0.74	1.12	29.07
	BC	0.38	0	0.3	0.4	1.01	0	0.76	1.02	3.87
	OC	0.34	0	1.11	0.14	0.9	0	2.79	0.37	5.65
	Reductions	40%	30%	0%	0%	40%	30%	0%	0%	–

G20. Emissions of NO<sub>x</sub> (NMVOCs) in CRA, Zhejiang province, and YRD were reduced by 69.3% (84.3%), 45.3% (56.3%), and 27.7% (37.5%), respectively, during G20.

### 2.3. Observations

The observed hourly concentrations of air pollutants are obtained from China's Ministry of Ecology and Environment and can be downloaded from [beijingair.sinaapp.com](http://beijingair.sinaapp.com). There are 11 observational sites in Hangzhou (Table 2), 10 of which are urban sites and one is a rural site. Daily mean values of PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, CO, NO<sub>2</sub>, and MDAS O<sub>3</sub> concentrations are calculated and used for model evaluation. The observed hourly meteorological parameters (temperature, relative humidity, wind, precipitation) are from Mantoushan monitoring station which is a national reference climate station.

**Table 2**  
Locations of observational sites in Hangzhou.

Sites	Longitude (°E)	Latitude (°N)
Binjiang <sup>a</sup>	120.22	30.21
Xixi <sup>a</sup>	120.06	30.27
Qiantao Lake <sup>b</sup>	119.03	29.64
Xiasha <sup>a</sup>	120.38	30.31
Wolong Bridge <sup>a</sup>	120.13	30.25
Zhejiang Agricultural University <sup>a</sup>	120.19	30.27
Zhaohui Wuqu <sup>a</sup>	120.16	30.29
Hemu Primary School <sup>a</sup>	120.12	30.31
Linping Town <sup>a</sup>	120.30	30.42
Chengxiang Town <sup>a</sup>	120.27	30.18
Yunxi <sup>a</sup>	120.09	30.18

<sup>a</sup> Urban sites.

<sup>b</sup> Rural sites.

### 2.4. Numerical experiments

The simulations of air pollutants are carried out for the time period of 10 August to 20 September, which are further classified as the periods of before G20 (PREG20, 10–23 August), during G20 (24 August to 6 September), and after G20 (POSTG20, 7–20 September). To compare the effects of local and joint emission controls, four numerical experiments are conducted (Table 3):

- (1) BASE: Baseline simulation without emission control measures from 10 August to 20 September;
- (2) CTRL\_CRA: Same as BASE simulation but with emission control in CRA during G20;
- (3) CTRL\_ZJ: Same as BASE simulation but with emission control over Zhejiang province during G20;
- (4) CTRL\_YRD: Same as BASE simulation but with emission control in YRD during G20;

Four more numerical experiments are conducted to quantify the sensitivity of O<sub>3</sub> concentrations to emission reductions in different sectors from 24 August to 6 September (Table 3):

- (5) CTRL\_noIND: Same as BASE simulation but no emissions from industry sector in YRD;
- (6) CTRL\_noPOW: Same as BASE simulation but no emissions from power sector in YRD;
- (7) CTRL\_noRES: Same as BASE simulation but no emissions from residential sector in YRD;
- (8) CTRL\_noTRA: Same as BASE simulation but no emissions from transportation sector in YRD.

All the simulations are driven by the assimilated MERRA-2 meteorological fields and there is a three-month spin-up before August 24 of 2016.



**Table 3**  
Summary of numerical experiments.

Simulation	Simulated time period	Sectors with emission control	Area with emission control
BASE	10 August–20 September	Without control	Without control
CTRL_CRA	24 August–6 September	Industry, Power, Residential, Transport	CRA <sup>a</sup>
CTRL_ZJ	24 August–6 September	Industry, Power, Residential, Transport	Zhejiang province <sup>b</sup>
CTRL_YRD <sup>d</sup>	24 August–6 September	Industry, Power, Residential, Transport	YRD <sup>c</sup>
CTRL_noIND	24 August–6 September	Industry	YRD
CTRL_noPOW	24 August–6 September	Power	YRD
CTRL_noRES	24 August–6 September	Residential	YRD
CTRL_noTRA	24 August–6 September	Transport	YRD

<sup>a</sup> Red area in the right panel of Fig. 1.

<sup>b</sup> Red plus orange plus light yellow areas in the right panel of Fig. 1.

<sup>c</sup> All colored areas under emission control in the right panel of Fig. 1.

<sup>d</sup> Control measures in CTRL\_YRD simulation were actually conducted during G20 meeting.

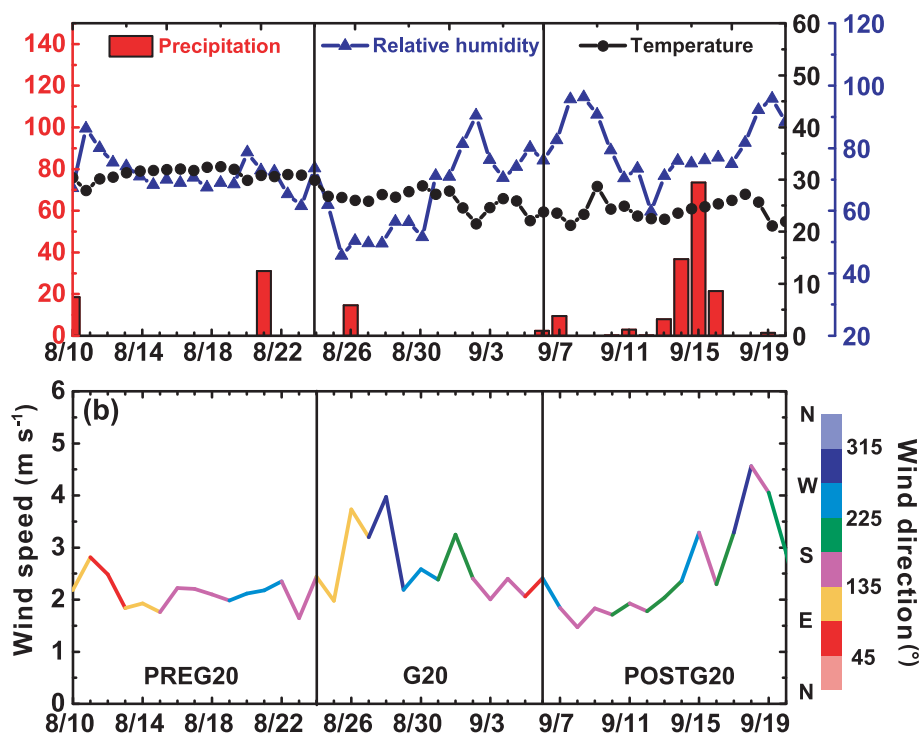
### 3. Results

#### 3.1. Observed meteorological parameters and concentrations of pollutants

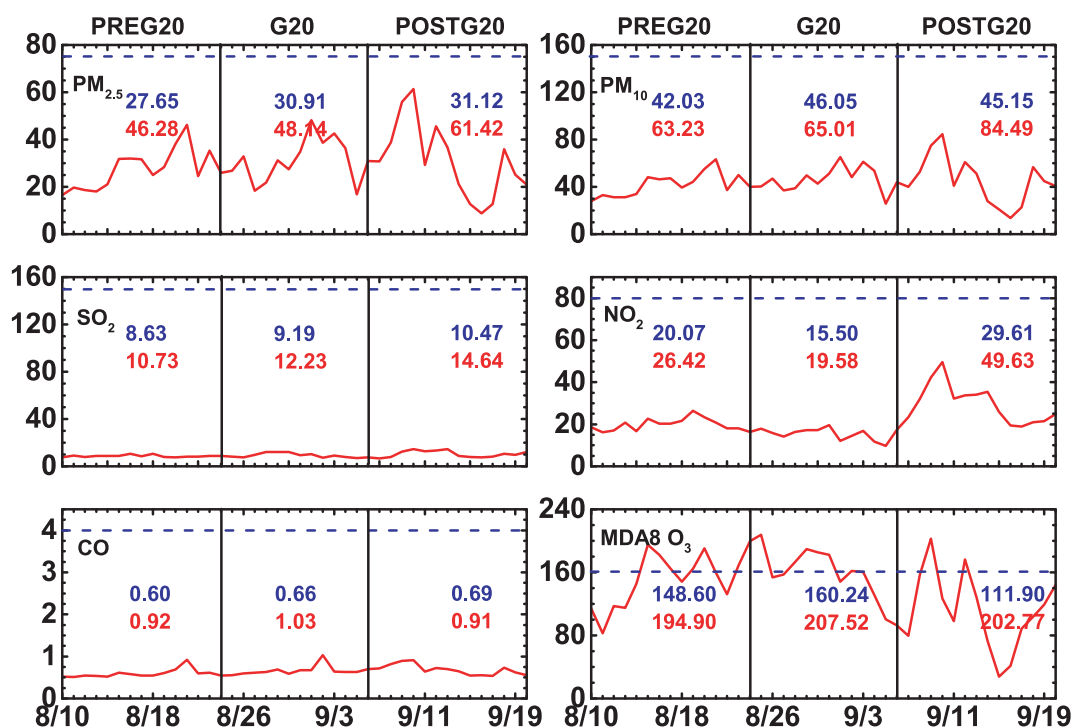
Fig. 3 shows the variations of temperature, precipitation, relative humidity, wind speed and wind direction in Hangzhou from 10 August to 20 September in 2016. During G20 (24 August to 6 September), mean temperature was 27.4 °C, which was 12.7% lower than that in PREG20 (31.4 °C) and 15.6% higher than the value in POSTG20 (23.7 °C). Averaged relative humidity during G20 was 60.4%, which was the lowest value among the three periods (it was 71.7% in PREG20 and 79.2% in POSTG20). The accumulated precipitation reached 49.5, 16.9, and 153.6 mm in PREG20, G20, and POSTG20, respectively. Daily wind speeds in the whole period (10 August–20 September) were generally low and most of them were lower than 4 m s<sup>-1</sup>. The mean wind speed during G20 was 2.6 m s<sup>-1</sup>, which was higher than that in PREG20 (2.1 m s<sup>-1</sup>). The

period of G20 had high temperature, low relative humidity, and low wind speed, which were favorable for O<sub>3</sub> formation as reported by Gong and Liao (2019) on the basis of the analysis of all observed O<sub>3</sub> pollution events in eastern China during 2014–2017.

Fig. 4 shows the observed daily surface concentrations of pollutants (PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub>, CO, MDA8 O<sub>3</sub>) in Hangzhou from 10 August to 20 September of 2016. During G20, concentrations of all pollutants, except for O<sub>3</sub>, were kept below national air quality standards. Relative to PREG20, regional mean concentrations of PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub>, CO, and MDA8 O<sub>3</sub> in Hangzhou (YRD) changed by +3.3 (+8.9) μg m<sup>-3</sup>, +4.0 (+16.7) μg m<sup>-3</sup>, +0.6 (+3.4) μg m<sup>-3</sup>, -4.6 (+6.8) μg m<sup>-3</sup>, +0.06 (+0.08) mg m<sup>-3</sup>, and +11.6 (+36.6) μg m<sup>-3</sup>, respectively (Fig. S1). Therefore, concentrations of all species increased over the YRD and the magnitudes of increases were larger than those of increases in Hangzhou. Concentration of NO<sub>2</sub> in Hangzhou showed decrease relative to PREG20. These changes indicate the effectiveness of control measures in Hangzhou. The overall increases in concentrations in YRD during G20



**Fig. 3.** (a) Time series of observed daily precipitation (red bars), relative humidity (blue line with triangles), temperature (black line with dots) at Mantoushan station in Hangzhou from 10 August to 20 September of 2016. (b) Time series of observed daily wind. The values are wind speeds and the colors show wind directions. (For interpretation of the references to color in this figure legend, the reader is referred to the Web version of this article.)



**Fig. 4.** Observed daily surface concentrations (red lines) of PM<sub>2.5</sub>, PM<sub>10</sub>, SO<sub>2</sub>, NO<sub>2</sub>, CO and MDA8 O<sub>3</sub> averaged over 11 sites in Hangzhou (Table 2) during 10 August to 20 September, 2016. The numbers indicate mean (blue) and maximum (red) concentrations of pollutants before, during and after G20 meeting. Units are mg m<sup>-3</sup> for CO and μg m<sup>-3</sup> for others. The blue dashed line in each panel represents national standard for the pollutant (75 μg m<sup>-3</sup> for PM<sub>2.5</sub>, 150 μg m<sup>-3</sup> for PM<sub>10</sub>, 150 μg m<sup>-3</sup> for SO<sub>2</sub>, 80 μg m<sup>-3</sup> for NO<sub>2</sub>, 4 mg m<sup>-3</sup> for CO, and 160 μg m<sup>-3</sup> for MDA8 O<sub>3</sub>). (For interpretation of the references to color in this figure legend, the reader is referred to the Web version of this article.)

relative to PREG20 were caused by meteorological conditions. As shown in Fig. S2, southeasterlies in PREG20 brought fresh air mass to YRD, while northerlies during G20 carried pollutants to YRD from polluted areas such as the North China Plain (NCP) and Henan province (Ni et al., 2018b; Zhang et al., 2020). The observed concentrations of all pollutants increased in the beginning of POSTG20. Therefore, the scientific question is whether the control measures reduced O<sub>3</sub> concentrations during the meeting or not.

### 3.2. Evaluation of model performance

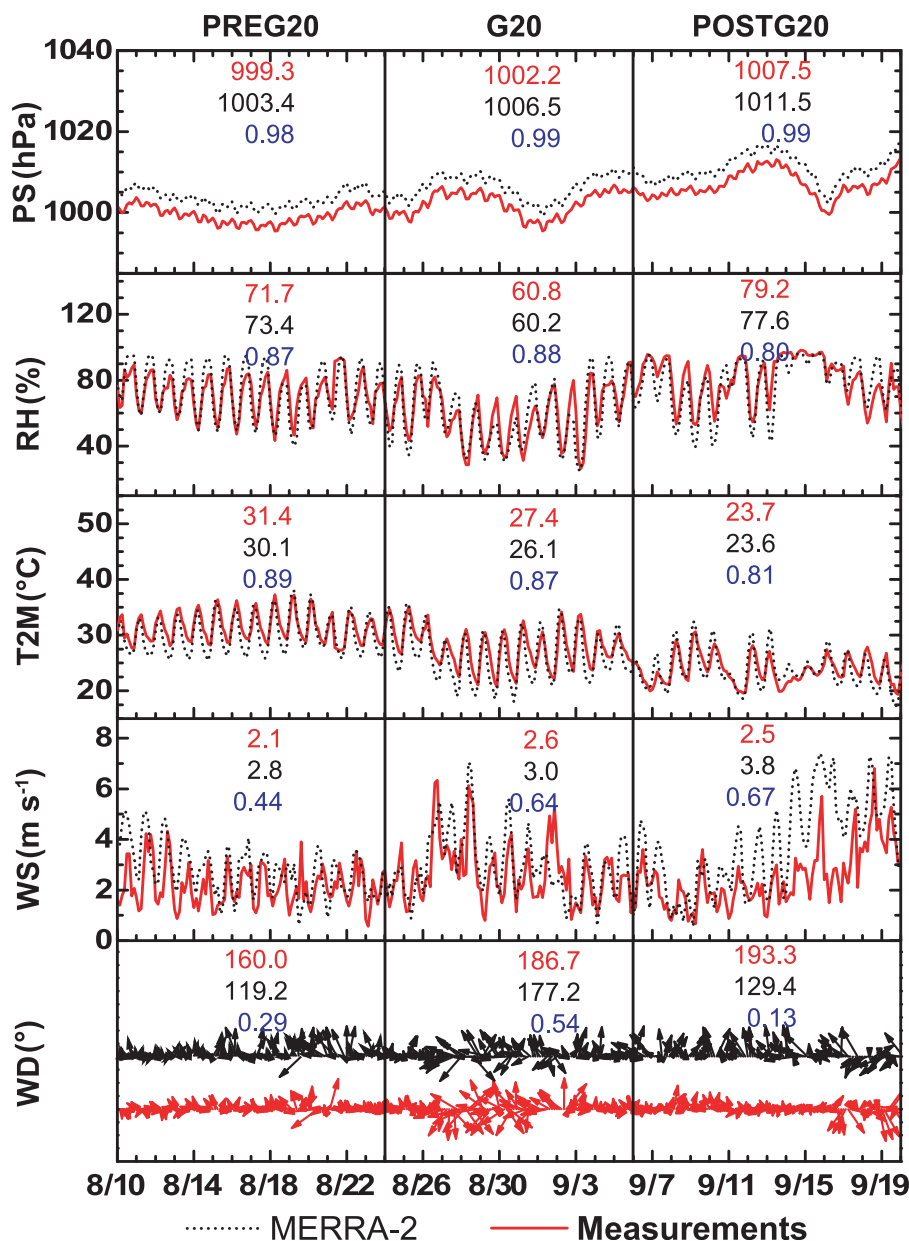
#### 3.2.1. Evaluation of MERRA-2 meteorological parameters

Fig. 5 shows the comparisons of MERRA-2 meteorological parameters with observations at the location of Hangzhou (119.8°E, 30.4°N) for sea-level pressure (SLP), relative humidity (RH), 2-m temperature (T2M), 10-m wind speed (WS) and 10-m wind direction from 10 August to 20 September of 2016. The observed parameters are averaged every 3 h to be consistent with the 3-h time resolution of MERRA-2. During 10 August to 20 September, MERRA-2 meteorological fields agree closely with measurements, with normalized mean biases (NMB =  $\frac{\sum_{i=1}^n (M_i - O_i)}{\sum_{i=1}^n O_i} \times 100\%$ ;  $M_i$  and  $O_i$  are the MERRA-2 (simulated) value and observed value of a meteorological parameter (pollutant concentration) at time (or site)  $i$ , respectively, and  $n$  is total number of samples) of 4.0%, -0.2%, -3.3%, 5.0% and -52.0% for SLP, RH, T2M, WS, and WD, respectively. During G20, the correlations between MERRA-2 and observed meteorological parameters were in the range of 0.54–0.99, indicating that the MERRA-2 meteorological parameters in Hangzhou agreed reasonably well with observations.

#### 3.2.2. Evaluation of simulated MDA8 O<sub>3</sub> from the CTRL\_YRD simulation

Fig. 6a shows the comparison of simulated and observed MDA8 O<sub>3</sub> concentrations in Hangzhou from 10 August to 20 September. Simulated magnitude and variation in MDA8 O<sub>3</sub> concentrations agree closely with observations in Hangzhou. In the periods of PREG20 (simulated by BASE), G20 (simulated by CTRL\_YRD), and POSTG20 (simulated by BASE), MDA8 O<sub>3</sub> concentrations have NMBs of 12.3%, -8.0%, and 31.1%, respectively. The correlation coefficients of MDA8 O<sub>3</sub> concentrations between BASE and observations are 0.70, 0.81, and 0.80 in these three time periods, respectively. Fig. 6b and c compare the spatial distributions of MDA8 O<sub>3</sub> concentrations in YRD simulated by CTRL\_YRD with observations for 25 August (the second day of G20 period) and 6 September (the final day of G20). On 25 August, observed and simulated MDA8 O<sub>3</sub> concentrations were high (exceeding 150 μg m<sup>-3</sup>) over a large area that covered provinces of Anhui, Jiangsu, and northern Zhejiang. The model captures fairly well the high MDA8 O<sub>3</sub> concentrations on 25 August (with  $r$  of 0.43 and NMB of -8.9%) and the low concentrations on 6 September (with  $r$  of 0.66 and NMB of 25.7%) considering all the observational sites in YRD.

The model performance shown here is similar to that of GEOS-Chem model reported in studies of Gu and Liao (2016) and Zhu et al. (2019). The correlation coefficients between observed and simulated pollutants (CO, NO<sub>2</sub>, SO<sub>2</sub>, and PM<sub>2.5</sub>) were in the range of 0.44–0.83 during APEC (Gu and Liao, 2016). The correlation coefficients (NMBs) between observed and simulated concentrations were 0.52 (-22.2%), 0.65 (+7.1%), 0.71 (-37.6%) for PM<sub>2.5</sub> and 0.91 (+25.0%), 0.86 (+31.9%), 0.85 (+20.7%) for MDA8 O<sub>3</sub> in Shanghai, Jiangyin, and Wenzhou, respectively, in July of 2016 (Zhu et al., 2019).

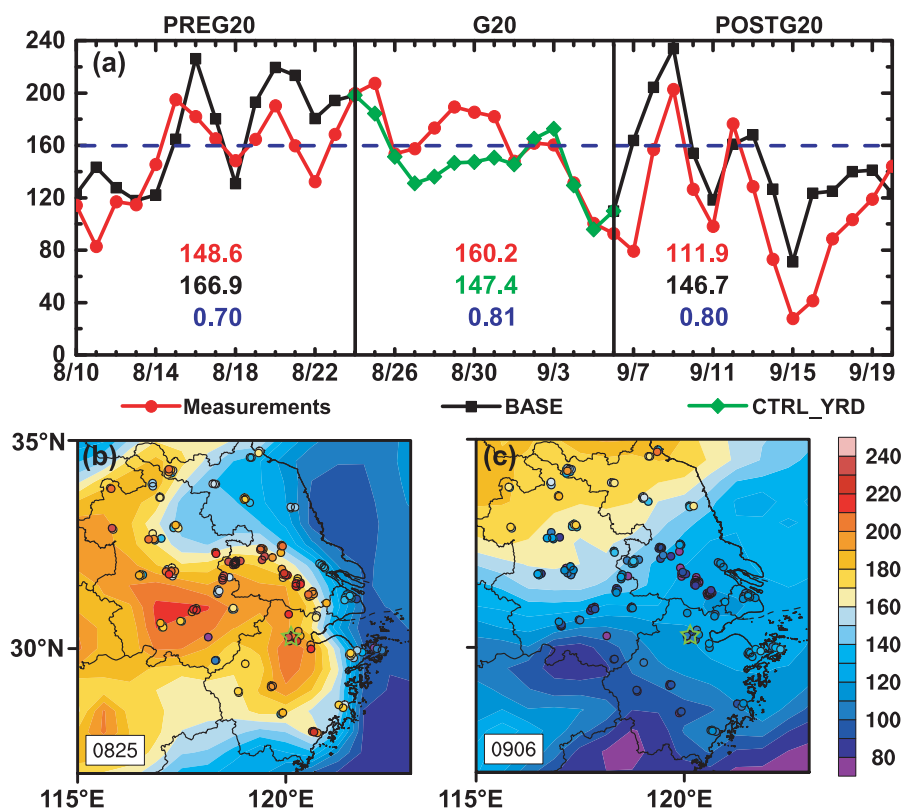


**Fig. 5.** Comparisons of MERRA-2 (black) and observed (red) sea-level pressure (SLP), relative humidity (RH), 2-m temperature (T2M), 10-m wind speed (WS) and 10-m wind direction (WD), at Mantoushan station in Hangzhou, from 10 August to 20 September of 2016. For each time period, the black, red, and blue numbers are the average of MERRA-2 variable, the average of observed value, and correlation coefficients (R) between measurements and MERRA-2 data, respectively. (For interpretation of the references to color in this figure legend, the reader is referred to the Web version of this article.)

### 3.3. Impact of district-joint control on $O_3$ concentrations during G20

Three numerical experiments (CTRL\_CRA, CTRL\_ZJ and CTRL\_YRD) are conducted to quantify the impacts of control measures on  $O_3$  during G20 as described in Section 2.4. Fig. 7a-c (Fig. 7d-f) show, relative to the BASE simulation, the absolute (percentage) changes in MDA8  $O_3$  concentrations resulting from emission controls in core area (CRA, red area in the right panel of Fig. 1), in the whole of Zhejiang Province (red plus orange plus light yellow areas in the right panel of Fig. 1), and in the whole of YRD (all colored areas in the right panel of Fig. 1), respectively, and concentrations are averaged over 24 August to 6 September. The largest reductions in MDA8  $O_3$  concentrations are simulated in northern

Zhejiang province. Compared to the BASE simulation, highest reductions in MDA8  $O_3$  concentrations are simulated to be,  $15.6 \mu\text{g m}^{-3}$  (8.0%),  $21.9 \mu\text{g m}^{-3}$  (11.7%), and  $26.0 \mu\text{g m}^{-3}$  (14.0%) in simulations of CTRL\_CRA, CTRL\_ZJ, and CTRL\_YRD, respectively. It should be noted that the location of highest reduction in MDA8  $O_3$  is simulated to be south of the core area, as a result of the high MDA8  $O_3$  concentrations in that location simulated by BASE because of high emissions of biogenic VOCs (Li et al., 2016b) and the prevailing northerlies at 850 hPa during G20 (Fig. S3a). At Hangzhou, averaged over 24 August to 6 September, while MDA8  $O_3$  concentration was  $160.2 \mu\text{g m}^{-3}$  from observations, simulated values are 166.9, 155.6, 152.1, and  $147.4 \mu\text{g m}^{-3}$  in the BASE, CTRL\_CRA, CTRL\_ZJ, and CTRL\_YRD simulations (Fig. 7d), respectively. Control measures in CRA, Zhejiang province, and YRD are



**Fig. 6.** (a) Comparison of daily MDA8 O<sub>3</sub> concentrations ( $\mu\text{g m}^{-3}$ ) from simulations (BASE: black line with squares; CTRL\_YRD: green line with diamonds) with observations (red line with dots) in Hangzhou from 10 August to 20 September of 2016. Blue dashed line represents national standard for O<sub>3</sub>. For each period of PREG20, G20, and POSTG20, mean concentration ( $\mu\text{g m}^{-3}$ ; red, black, and green number for observations, BASE and CTRL\_YRD simulations, respectively) of MDA8 O<sub>3</sub> as well as the correlation coefficient between simulated and observed MDA8 O<sub>3</sub> concentrations (blue) are indicated. Spatial distributions of simulated (shades) by CTRL\_YRD and observed (dots) concentrations of MDA8 O<sub>3</sub> ( $\mu\text{g m}^{-3}$ ) in YRD are shown for (b) 25 August and (c) 6 September. The star indicates the location of Hangzhou. (For interpretation of the references to color in this figure legend, the reader is referred to the Web version of this article.)

simulated to lead to, respectively, 6.8%, 8.9%, and 11.7% of reductions in averaged MDA8 O<sub>3</sub> in Hangzhou during G20 relative to the BASE simulation. These results suggest that the emission control measures reduced O<sub>3</sub> concentrations in Hangzhou effectively although O<sub>3</sub> concentrations with emission controls still exceeded the national standard of  $160 \mu\text{g m}^{-3}$  during G20. The reductions in O<sub>3</sub> concentrations with emission controls in YRD are higher than those with controls in core areas or in the whole of Zhejiang province, indicating the effectiveness of district-joint emission control measures. These model results also suggest that, in reality, regions to carry out emission control can be decided on the basis of the targeted magnitude of reduction in O<sub>3</sub>, since emission control in the core area alone improved O<sub>3</sub> pollution in Hangzhou.

#### 3.4. VOCs-NO<sub>x</sub>-O<sub>3</sub> sensitivity during G20

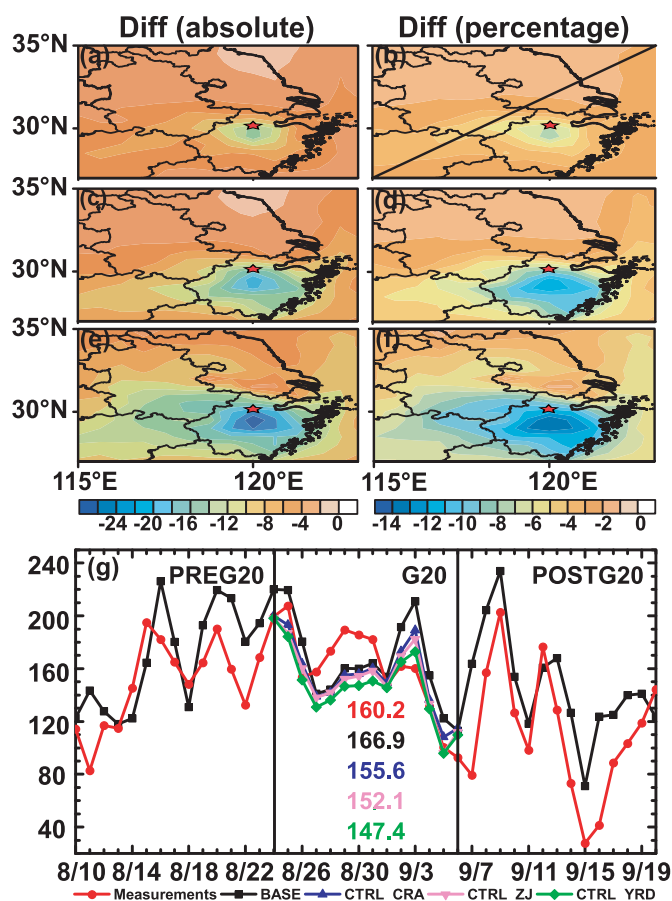
As the most important precursors of O<sub>3</sub>, NO<sub>x</sub> (NO + NO<sub>2</sub>) and VOCs play crucial roles in the formation of tropospheric O<sub>3</sub>, and controlling them properly can reduce O<sub>3</sub> concentration effectively (Rabl and Eyre, 1998; Atkinson, 2000; Shao et al., 2009; Wei et al., 2014). O<sub>3</sub> formation is a highly nonlinear process. The O<sub>3</sub>-precursors relationship can be classified into the NO<sub>x</sub>-limited, VOC-limited, and transitional regimes quantified by VOCs/NO<sub>x</sub> ratios (Sillman, 1999; Sillman and He, 2002; Ran et al., 2009; Prabamroong et al., 2012; Strong et al., 2013; Zou et al., 2015; Jia et al., 2016; Li et al., 2017). In the VOC-limited (NO<sub>x</sub>-limited) regime, O<sub>3</sub> decreases with decreasing VOCs (NO<sub>x</sub>) and increases with decreasing NO<sub>x</sub> (VOCs). In transition regime, O<sub>3</sub> is sensitive to

both NO<sub>x</sub> and VOCs (Milford et al., 1989; Seinfeld and Pandis, 2006). We used the typical VOCs/NO<sub>x</sub> ratios which were applied by empirical kinetic modeling approach (EKMA) to classify sensitivity regimes during G20 to indicate the possible O<sub>3</sub> response to changes in VOCs or NO<sub>x</sub> concentrations. O<sub>3</sub> formation can be VOC-limited if the ratios are less than 4 and can be NO<sub>x</sub>-limited if the ratios are larger than 15. The VOCs/NO<sub>x</sub> ratios of 4–15 indicate a transitional regime (Lou et al., 2010; Edson et al., 2017; Li et al., 2017a), with which O<sub>3</sub> concentrations will increase with increasing NO<sub>x</sub> and VOCs and vice versa (Sillman, 1999).

Fig. 9 shows the simulated distribution of the ratios of VOCs/NO<sub>x</sub> in YRD from the GEOS-Chem model averaged over 24 August to 6 September. Following Lou et al. (2010), VOCs considered include ethane, propane, alkanes with more than 4 carbon atoms, propene, ketones, Methyl ethyl ketone, isoprene, other aldehydes, aldehydes with more than 3 carbon atoms, methyl ketone, methacrolein, and formaldehyde. VOCs/NO<sub>x</sub> ratios are simulated to be 9.9 in the core area, 17.7 in southern Zhejiang province, and 3.8 in southern Jiangsu province. Correspondingly, central and northern Anhui and southern Jiangsu were VOC-limited, western and southwestern Zhejiang were NO<sub>x</sub>-limited, and northern, central (core area), and southeastern Zhejiang were in the transitional regime. The regimes we identified agrees with the results of previous studies using models and observations (Jin and Holloway, 2015; Li et al., 2017a; Wang et al., 2019). Therefore, reductions in either NO<sub>x</sub> or VOCs reduced O<sub>3</sub> concentrations effectively in Hangzhou during G20.

It should be noted that the VOCs/NO<sub>x</sub> ratio can only be considered as an indicator to analyze the possible ozone response to





**Fig. 7.** (a)–(c) Absolute differences ( $\mu\text{g m}^{-3}$ ) and (d)–(f) percentage changes (%) in MDA8  $\text{O}_3$  concentrations in CTRL\_CRA, CTRL\_ZJ, and CTRL\_YRD simulations (from top to bottom) relative to the BASE simulation. (g) Comparisons of daily MDA8  $\text{O}_3$  concentrations ( $\mu\text{g m}^{-3}$ ) from the BASE (black line), CTRL\_CRA (blue line), CTRL\_ZJ (cyan line) and CTRL\_YRD (green line) simulations with observations (red line) in Hangzhou from 10 August to 20 September of 2016. The colored numbers are the averaged MDA8  $\text{O}_3$  concentrations ( $\mu\text{g m}^{-3}$ ) during G20 from measurements and simulations. The red star in (a)–(f) indicates the location of Hangzhou. (For interpretation of the references to color in this figure legend, the reader is referred to the Web version of this article.)

changes in VOCs or  $\text{NO}_x$  concentrations because this rule does not account for the impact of VOCs reactivity, biogenic response, geographic variations (Sillman, 1999). The typical values for differentiating the two regimes can also be different according to regions and time periods in previous studies (Milford et al., 1994; Hanna et al., 1996; Zou et al., 2015).

### 3.5. Impact of emission reductions by sectors on $\text{O}_3$

According to the measurements during G20, mean concentration of MDA8  $\text{O}_3$  in Hangzhou ( $160.2 \mu\text{g m}^{-3}$ ) was above the national air quality standard (Fig. 4). Extra emission reductions should be carried out to keep  $\text{O}_3$  concentrations below national standard. It will be helpful to examine the sensitivities of  $\text{O}_3$  concentrations in Hangzhou to emissions in different sectors (industry, power, transportation, and residential sectors) during G20. Four numerical experiments are conducted as described in Section 2.4.

Fig. 9 shows, relative to the BASE simulation, the changes in mean MDA8  $\text{O}_3$  concentrations in Hangzhou during G20 when emissions of a specific sector are set to zero in YRD. The emissions from industry sector play a dominant role in  $\text{O}_3$  formation in YRD. The removal of emissions of all anthropogenic chemical species ( $\text{NO}_x$ , VOCs, CO, BC, OC,  $\text{NH}_3$ ,  $\text{SO}_2$ ) from industry sector would

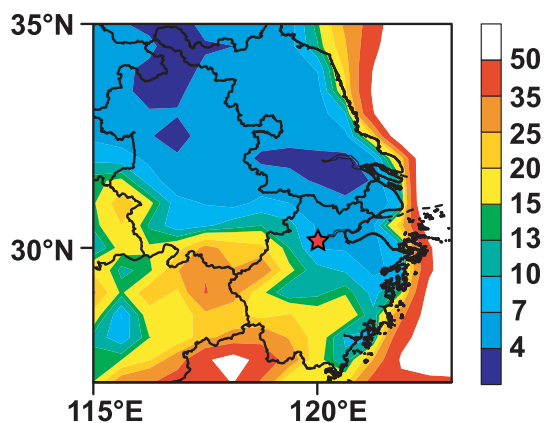
reduce MDA8  $\text{O}_3$  concentrations in the whole of YRD especially in Zhejiang province during G20 (Fig. 9a), with the highest reductions of  $23.1 \mu\text{g m}^{-3}$  in MDA8  $\text{O}_3$  concentrations. Simulated distribution of reductions in MDA8  $\text{O}_3$  concentrations is similar to that of simulated MDA8  $\text{O}_3$  in the BASE simulation (Fig. S3b). The emissions of both  $\text{NO}_x$  and VOCs from industry sector are high in southern Jiangsu province, Shanghai, and northern Zhejiang province (Figs. S4a and S4e). The largest reductions in MDA8  $\text{O}_3$  in central Zhejiang in Fig. 9a is a result of the transitional regime in this region (Fig. 8). Concentrations of MDA8  $\text{O}_3$  in Hangzhou during G20 would decrease by  $17.6 \mu\text{g m}^{-3}$  (10.5%) due to cutting off emissions from industry sector in YRD.

In power sector, emissions of  $\text{NO}_x$  are dominant and those of VOCs are practically zero. The highest  $\text{NO}_x$  emissions are located over Southern Jiangsu and Shanghai (Figs. S4b and S4f). The removal of emissions from power sector would have an adverse effect of increasing MDA8  $\text{O}_3$  concentrations by  $0\text{--}6 \mu\text{g m}^{-3}$  in Southern Jiangsu and Shanghai (Fig. 9b) where VOCs/ $\text{NO}_x$  ratios exhibit a VOC-limited regime. In Hangzhou, the removal of emissions from power sector in YRD is simulated to reduce MDA8  $\text{O}_3$  concentrations during G20 by  $5.1 \mu\text{g m}^{-3}$  (3.1%).

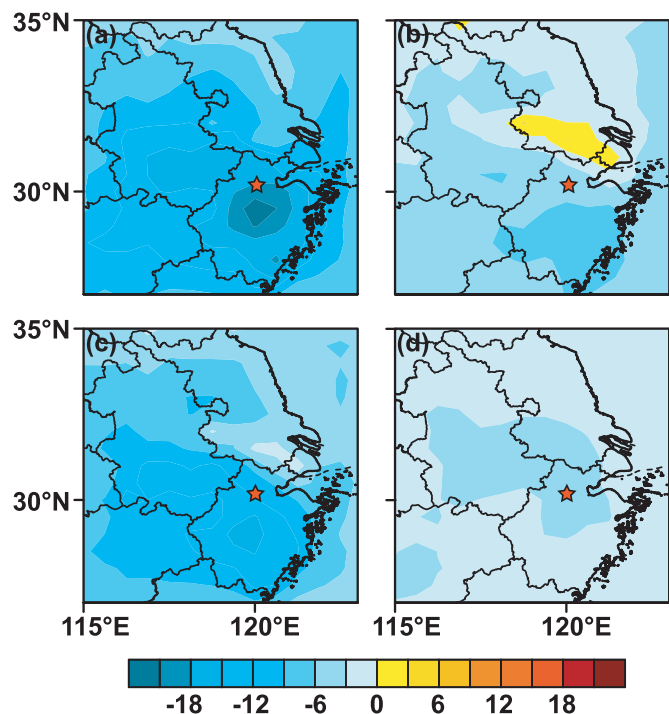
In transportation sector, emissions of  $\text{NO}_x$  are much higher than those of VOCs (Figs. S2e and S2g), and emissions are generally high in southern Jiangsu, Shanghai, and Hangzhou. The removal of emissions from transportation sector would reduce concentrations of MDA8  $\text{O}_3$  in the whole of YRD, especially in central and southern Zhejiang province where highest reductions of about  $16.0 \mu\text{g m}^{-3}$  are simulated. Over southern Jiangsu and Shanghai where  $\text{O}_3$  formation is VOC-limited, the reductions in MDA8  $\text{O}_3$  are small (less than  $3 \mu\text{g m}^{-3}$ ). Without emissions from transportation sector, concentrations of MDA8  $\text{O}_3$  in Hangzhou during G20 would be reduced by  $12.3 \mu\text{g m}^{-3}$  (7.4%) relative to the BASE simulation.

Among the sectors considered here, residential sector has the lowest emissions of  $\text{NO}_x$  and the second lowest emissions of VOCs (Figs. S2d and S2h). Emissions of VOCs are larger than those of  $\text{NO}_x$ , so the removal of emissions from residential sector would reduce MDA8  $\text{O}_3$  concentrations in the whole of YRD. Reductions in MDA8  $\text{O}_3$  concentrations of  $3\text{--}6 \mu\text{g m}^{-3}$  are simulated in northern Zhejiang, southern Jiangsu, and southern Anhui provinces. No emissions from residential sector would reduce concentrations of MDA8  $\text{O}_3$  in Hangzhou by  $3.2 \mu\text{g m}^{-3}$  (1.9%) during G20.

In summary, to further reduce  $\text{O}_3$  in Hangzhou, reductions in emissions from industry and transportation sectors can be the most effective. Reductions in industry sector over the strictly controlled



**Fig. 8.** Simulated distribution of the ratios of VOCs/ $\text{NO}_x$  averaged over 24 August to 6 September. The red star indicates the location of Hangzhou. (For interpretation of the references to color in this figure legend, the reader is referred to the Web version of this article.)



**Fig. 9.** Simulated mean changes in MDA8 O<sub>3</sub> concentrations ( $\mu\text{g m}^{-3}$ ) during G20 in simulations (a) without emissions from industry sector (CTRL\_noIND), (b) without emissions from power sector (CTRL\_noPOW), (c) without emissions from transportation sector (CTRL\_noTRA), and (d) without emissions from residential sector (CTRL\_noRES) in YRD relative to BASE simulation. The red star in (a)–(d) indicates the location of Hangzhou. (For interpretation of the references to color in this figure legend, the reader is referred to the Web version of this article.)

area (Fig. 1) would be especially helpful for keeping the MDA8 O<sub>3</sub> concentrations in Hangzhou during G20 below the national standard considering the transitional regime in that region. It should be noted that the simulations with 100% emission reductions in each sector represent an ideal assumption to evaluate the impact of emission reductions by sectors on O<sub>3</sub>. The percentage of emission reductions conducted in core area in Hangzhou of different sectors were in the range of 50%–100% during G20 (Table 1), so the actual impacts of the emission reductions conducted for the four sectors would be less than those obtained from the sensitivity studies.

#### 4. Conclusion

Summertime O<sub>3</sub> concentrations have been increasing over eastern China since 2013. Here we take G20 meeting as an example to investigate whether and how the emission control measures can reduce O<sub>3</sub> concentrations under the typical weather condition during late summer to early fall. The emissions, concentrations, and meteorological conditions are examined from 10 August to 20 September in 2016, covering the periods of PREG20, G20, and POSTG20. The nested-grid version of the GEOS-Chem model is used to assess the effectiveness of emission control measures carried out during G20.

During G20 meeting, observed concentrations of all air pollutants were below national air quality standards except for O<sub>3</sub>. With the emission control measures, mean MDA8 O<sub>3</sub> concentration was  $160.2 \mu\text{g m}^{-3}$  during G20. Analyses of meteorological parameters show that the period of G20 had high temperature ( $27.4 \text{ }^\circ\text{C}$ ), low relative humidity (60.4%), low wind speed ( $2.6 \text{ m s}^{-1}$ ) and insufficient precipitation (16.9 mm), all of which were favorable for O<sub>3</sub>

formation.

Model evaluation shows that the GEOS-Chem model can capture well the daily variations of MDA8 O<sub>3</sub> in Hangzhou, with correlation coefficients (NMBs) of 0.71 (12.3%), 0.81 (−8.0%), and 0.80 (31.1%) in PREG20, G20, and POSTG20, respectively. Model simulations show that, during G20, the control measures in the core control area, Zhejiang province, and the whole of YRD lead to reductions in MDA8 O<sub>3</sub> in Hangzhou by 6.8%, 8.9%, and 11.7%, respectively, indicating that it is essential to carry out district-joint control measures during G20. Ratios of VOCs/NO<sub>x</sub> in YRD during G20 indicate that the core area and most of Zhejiang province were in transitional zone ( $4 \leq \text{VOCs/NO}_x \leq 15$ ); reductions in either NO<sub>x</sub> or VOCs can reduce concentrations of O<sub>3</sub> effectively.

Sensitivities of MDA8 O<sub>3</sub> concentrations to emission reductions in different sectors, including industry, power, transportation, and residential sectors, are also examined. Removal of emissions of all chemical species from these sectors are simulated to reduce the MDA8 O<sub>3</sub> concentrations in Hangzhou during G20 by  $17.6 \mu\text{g m}^{-3}$  (10.5%),  $5.1 \mu\text{g m}^{-3}$  (3.1%),  $12.3 \mu\text{g m}^{-3}$  (7.4%), and  $3.2 \mu\text{g m}^{-3}$  (1.9%), respectively. Therefore, the control of emissions from industry and transportation sectors in YRD can be the most effective to reduce O<sub>3</sub> in the YRD. Considering the location and the NO<sub>x</sub>/VOCs ratios, further reductions in industry sector in the strictly controlled area would be especially helpful to keep concentrations of O<sub>3</sub> in Hangzhou below the national standard during G20.

On the basis of our results, we argue that O<sub>3</sub> pollution can be controlled in summer by emission reductions under the background of seriously increasingly O<sub>3</sub> pollution in China. The effectiveness of emission control measures in Hangzhou depend on the local transitional regime of VOCs/NO<sub>x</sub> ratios in summer, which may not be the situation in other places of eastern China. Our model results may also have uncertainties in emissions inventories and the assumed reduction rates. This work highlights the necessity of district-joint control and the sensitivities of O<sub>3</sub> to reductions in different sectors provide suggestions for O<sub>3</sub> control in YRD in the future.

#### Author contribution

HL and YW conceived the study and designed the experiments. YW carried out the simulations and performed the analysis. YW and HL prepared the manuscript.

#### Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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#### Appendix A. Supplementary data

Supplementary data to this article can be found online at <https://doi.org/10.1016/j.chemosphere.2020.127729>.

#### References

- Alexander, B., Park, R.J., Jacob, D.J., Li, Q., Yantosca, R.M., Savarino, J., Lee, C.C., Thiemens, M.H., 2005. Sulfate formation in sea-salt aerosols: constraints from

- oxygen isotopes. *J. Geophys. Res.* 110 <https://doi.org/10.1029/2004JD005659>.
- Atkinson, R., 1998. Atmospheric chemistry of VOCs and NO<sub>x</sub>. *Atmos. Environ.* 34, 2063–2101. [https://doi.org/10.1016/S1352-2310\(99\)00460-4](https://doi.org/10.1016/S1352-2310(99)00460-4).
- Bey, I., Jacob, D.J., Yantosca, R.M., Logan, J.A., Field, B.D., Fiore, A.M., Li, Q., Liu, H.Y., Mickley, L.J., Schultz, M.G., 2001. Global modeling of tropospheric chemistry with assimilated meteorology: model description and evaluation. *J. Geophys. Res.* 106, 23073–23095. <https://doi.org/10.1029/2001JD000807>.
- Bian, H., Prather, M.J., 2002. Fast-J2: accurate simulation of stratospheric photolysis in global chemical models. *J. Atmos. Chem.* 41, 281–296. <https://doi.org/10.1023/A:1014980619462>.
- Cai, W., Li, K., Liao, H., Wang, H., Wu, L., 2017. Weather conditions conducive to Beijing severe haze more frequent under climate change. *Nat. Clim. Change* 7, 257–262. <https://doi.org/10.1038/nclimate3249>.
- Edson, C.T., Iván, H.-P., Alberto, M., 2017. Use of combined observational- and model-derived photochemical indicators to assess the O<sub>3</sub>-NO<sub>x</sub>-VOC System sensitivity in urban areas. *Atmosphere* 8, 22. <https://doi.org/10.3390/atmos8020022>.
- Evans, M.J., Jacob, D.J., 2005. Impact of new laboratory studies of N<sub>2</sub>O<sub>5</sub> hydrolysis on global model budgets of tropospheric nitrogen oxides, ozone, and OH. *Geophys. Res. Lett.* 32 <https://doi.org/10.1029/2005GL022469>.
- Fairlie, T.D., Jacob, D.J., Park, R.J., 2007. The impact of transpacific transport of mineral dust in the United States. *Atmos. Environ.* 41, 1251–1266. <https://doi.org/10.1016/j.atmosenv.2006.09.048>.
- Fu, J.S., Streets, D.G., Jang, C., Hao, J., He, K., Wang, L., Zhang, Q., 2009. Modeling regional/urban ozone and particulate matter in Beijing, China. *J. Air Waste Manag. Assoc.* 59, 37–44. <https://doi.org/10.3155/1047-3289.59.1.37>.
- Gao, Y., Liu, X., Zhao, C., Zhang, M., 2011. Emission controls versus meteorological conditions in determining aerosol concentrations in Beijing during the 2008 Olympic Games. *Atmos. Chem. Phys.* 11, 12437–12451. <https://doi.org/10.5194/acp-11-12437-2011>.
- Gong, C., Liao, H., 2019. A typical weather pattern for ozone pollution events in North China. *Atmos. Chem. Phys.* 19, 13725–13740. <https://doi.org/10.5194/acp-19-13725-2019>.
- Gu, Y., Liao, H., 2016. Response of fine particulate matter to reductions in anthropogenic emissions in Beijing during the 2014 Asia–Pacific Economic Cooperation summit. *Atmospheric Ocean. Sci. Lett.* 9, 411–419. <https://doi.org/10.1080/16742834.2016.1230465>.
- Guo, J., He, J., Liu, H., Miao, Y., Liu, H., Zhai, P., 2016. Impact of various emission control schemes on air quality using WRF-Chem during APEC China 2014. *Atmos. Environ.* 140, 311–319. <https://doi.org/10.1016/j.atmosenv.2016.05.046>.
- Hanna, S.R., Moore, G.E., Fernau, M.E., 1996. Evaluation of photochemical grid models (UAM-IV, UAM-V, and the ROM/UAM-IV couple) using data from the lake Michigan ozone study (LMOS). *Atmos. Environ.* 30, 3265–3279. [https://doi.org/10.1016/1352-2310\(96\)00060-X](https://doi.org/10.1016/1352-2310(96)00060-X).
- Huang, Q., Tijian, W., Pulong, C., Xiaoxian, H., Jialei, Z., Bingliang, Z., 2017. Impacts of emission reduction and meteorological conditions on air quality improvement during the 2014 Youth Olympic Games in Nanjing, China. *Atmos. Chem. Phys.* 1–19 <https://doi.org/10.5194/acp-17-13457-2017>.
- Huang, R., Zhang, Y., Bozzetti, C., Ho, K., Cao, J., Han, Y., Daellenbach, K.R., Slowik, J.G., Platt, S.M., Canonaco, F., 2014. High secondary aerosol contribution to particulate pollution during haze events in China. *Nature* 514, 218–222. <https://doi.org/10.1038/nature13774>.
- Jacob, D.J., 2000. Heterogeneous chemistry and tropospheric ozone. *Atmos. Environ.* 34, 2131–2159. [https://doi.org/10.1016/S1352-2310\(99\)00462-8](https://doi.org/10.1016/S1352-2310(99)00462-8).
- Jia, C., Mao, X., Huang, T., Liang, X., Wang, Y., Shen, Y., Jiang, W., Wang, H., Bai, Z., Ma, M., 2016. Non-methane hydrocarbons (NMHCs) and their contribution to ozone formation potential in a petrochemical industrialized city, Northwest China. *Atmos. Res.* 169, 225–236. <https://doi.org/10.1016/j.atmosres.2015.10.006>.
- Jin, X., Holloway, T., 2015. Spatial and temporal variability of ozone sensitivity over China observed from the Ozone Monitoring Instrument. *J. Geophys. Res.* 120, 7229–7246. <https://doi.org/10.1002/2015JD023250>.
- Li, K., Chen, L., Ying, F., White, S.J., Jang, C., Wu, X., Gao, X., Hong, S., Shen, J., Azzi, M., 2017a. Meteorological and chemical impacts on ozone formation: a case study in Hangzhou, China. *Atmos. Res.* 196, 40–52. <https://doi.org/10.1016/j.atmosres.2017.06.003>.
- Li, K., Liao, H., Zhu, J., Moch, J.M., 2016a. Implications of RCP emissions on future PM<sub>2.5</sub> air quality and direct radiative forcing over China. *J. Geophys. Res.* 121 <https://doi.org/10.1002/2016JD025623>.
- Li, L., An, J., Shi, Y.Y., Zhou, M., Yan, R., Huang, C., Wang, H., Lou, S., Wang, Q., Lu, Q., 2016b. Source apportionment of surface ozone in the Yangtze River Delta, China in the summer of 2013. *Atmos. Environ.* 144, 194–207. <https://doi.org/10.1016/j.atmosenv.2016.08.076>.
- Li, P., Wang, L., Guo, P., Yu, S., Mehmood, K., Wang, S., Liu, W., Seinfeld, J.H., Zhang, Y., Wong, D.C., 2017b. High reduction of ozone and particulate matter during the 2016 G-20 summit in Hangzhou by forced emission controls of industry and traffic. *Environ. Chem. Lett.* 15, 709–715. <https://doi.org/10.1007/s10311-017-0642-2>.
- Li, X., Qiao, Y., Zhu, J., Shi, L., Wang, Y., 2017c. The “APEC blue” endeavor: causal effects of air pollution regulation on air quality in China. *J. Clean. Prod.* 168, 1381–1388. <https://doi.org/10.1016/j.jclepro.2017.08.164>.
- Liang, P., Zhu, T., Fang, Y., Li, Y.P., Han, Y., Wu, Y., Hu, M., Wang, J., 2017. The role of meteorological conditions and pollution control strategies in reducing air pollution in Beijing during APEC 2014 and Victory Parade 2015. *Atmos. Chem. Phys.* 17, 13921–13940. <https://doi.org/10.5194/acp-17-13921-2017>.
- Liu, H., Jacob, D.J., Bey, I., Yantosca, R.M., 2001. Constraints from <sup>210</sup>Pb and <sup>7</sup>Be on wet deposition and transport in a global three-dimensional chemical tracer model driven by assimilated meteorological fields. *J. Geophys. Res.* 106, 12109–12128. <https://doi.org/10.1029/2000JD900839>.
- Lou, S., Zhu, B., Liao, H., 2010. Impacts of O<sub>3</sub> precursor on surface O<sub>3</sub> concentration over China. *Trans. Atmos. Sci.* 451–459, 033.
- Mao, M., Hu, D., 2017. Evaluation of the air pollution control over Zhejiang province during the G20 summit in Hangzhou. *Res. Environ. Sci.* 30.
- Milford, J.B., Russell, A.G., Mcrae, G.J., 1989. A new approach to photochemical pollution control: implications of spatial patterns in pollutant responses to reductions in nitrogen oxides and reactive organic gas emissions. *Environ. Sci. Technol.* 23, 1290–1301. <https://doi.org/10.1021/es00068a017>.
- Milford, J.B., Gao, D., Sillman, S., Blossy, P.N., Russell, A.G., 1994. Total reactive nitrogen (NO<sub>y</sub>) as an indicator of the sensitivity of ozone to reductions in hydrocarbon and NO<sub>x</sub> emissions. *J. Geophys. Res.* 99, 3533–3542. <https://doi.org/10.1029/93JD03224>.
- Ni, R., Lin, J., Yan, Y., Lin, W., 2018a. Foreign and domestic contributions to spring-time ozone over China. *Atmos. Chem. Phys.* 18, 1–39. <https://doi.org/10.5194/acp-2017-1226>.
- Ni, Z., Luo, K., Gao, Y., Jiang, F., Gao, X., Fan, J., Chen, C., 2018b. Modeling tropospheric O<sub>3</sub> evolution during the 2016 Group of twenty summit in Hangzhou, China. *Atmos. Chem. Phys.* 1–25. <https://doi.org/10.5194/acp-2018-76>.
- Park, R.J., Jacob, D.J., Chin, M., Martin, R.V., 2003. Sources of carbonaceous aerosols over the United States and implications for natural visibility. *J. Geophys. Res.* 108 <https://doi.org/10.1029/2002JD003190>.
- Park, R.J., Jacob, D.J., Field, B.D., Yantosca, R.M., Chin, M., 2004. Natural and transboundary pollution influences on sulfate-nitrate-ammonium aerosols in the United States: implications for policy. *J. Geophys. Res.* 109 <https://doi.org/10.1029/2003JD004473>.
- Prabamroong, T., Manomaiphiboon, K., Limpaseni, W., Sukhapan, J., Bonnet, S., 2012. Ozone and its potential control strategy for Chon Buri city, Thailand. *J. Air Waste Manag. Assoc.* 62, 1411–1422. <https://doi.org/10.1080/10962247.2012.716385>.
- Pye, H.O.T., Liao, H., Wu, S., Mickley, L.J., Jacob, D.J., Henze, D.K., Seinfeld, J.H., 2009. Effect of changes in climate and emissions on future sulfate-nitrate-ammonium aerosol levels in the United States. *J. Geophys. Res.* 114 <https://doi.org/10.1029/2008JD010701>.
- Qiu, Y., Liao, H., Zhang, R., Hu, J., 2017. Simulated impacts of direct radiative effects of scattering and absorbing aerosols on surface layer aerosol concentrations in China during a heavily polluted event in February 2014. *J. Geophys. Res.* 122, 5955–5975. <https://doi.org/10.1002/2016JD026309>.
- Rabl, A., Eyre, N., 1998. An estimate of regional and global O<sub>3</sub> damage from precursor NO<sub>x</sub> and VOC emissions. *Environ. Int.* 24, 835–850. [https://doi.org/10.1016/S0160-4120\(98\)00053-1](https://doi.org/10.1016/S0160-4120(98)00053-1).
- Ran, L., Zhao, C., Geng, F., Tie, X., Tang, X., Peng, L., Zhou, G., Yu, Q., Xu, J., Guenther, A., 2009. Ozone photochemical production in urban Shanghai, China: analysis based on ground level observations. *J. Geophys. Res.* 114 <https://doi.org/10.1029/2008JD010752>.
- Seinfeld, J.H., Pandis, S.N., 2006. *Atmospheric Chemistry and Physics: from Air Pollution to Climate Change*, second ed. John Wiley & Sons, New York.
- Shao, M., Zhang, Y., Zeng, L., Tang, X., Zhang, J., Zhong, L., Wang, B., 2009. Ground-level ozone in the pearl River Delta and the roles of VOC and NO<sub>x</sub> in its production. *J. Environ. Manag.* 90, 512–518. <https://doi.org/10.1016/j.jenvman.2007.12.008>.
- Sillman, S., 1999. The relation between ozone, NO<sub>x</sub> and hydrocarbons in urban and polluted rural environments. *Atmos. Environ.* 33, 1821–1845. [https://doi.org/10.1016/S1352-2310\(98\)00345-8](https://doi.org/10.1016/S1352-2310(98)00345-8).
- Sillman, S., He, D., 2002. Some theoretical results concerning O<sub>3</sub>-NO<sub>x</sub>-VOC chemistry and NO<sub>x</sub>-VOC indicators. *J. Geophys. Res.* 107 <https://doi.org/10.1029/2001JD001123>.
- Strong, J., Whyatt, J.D., Metcalfe, S.E., Derwent, R.G., Hewitt, C.N., 2013. Investigating the impacts of anthropogenic and biogenic VOC emissions and elevated temperatures during the 2003 ozone episode in the UK. *Atmos. Environ.* 74, 393–401. <https://doi.org/10.1016/j.atmosenv.2013.04.006>.
- Su, W., Liu, C., Hu, Q., Fan, G., Xie, Z., Huang, X., Zhang, T., Chen, Z., Dong, Y., Ji, X., 2017. Characterization of ozone in the lower troposphere during the 2016 G20 conference in Hangzhou. *Sci. Rep.* 7 <https://doi.org/10.1038/s41598-017-17646-x>, 17368–17368.
- Tang, G., Zhu, X., Xin, J., Hu, B., Song, T., Sun, Y., Zhang, J., Wang, L., Cheng, M., Chao, N., 2017. Modelling study of boundary-layer ozone over northern China - Part I: ozone budget in summer. *Atmos. Res.* 187, 128–137. <https://doi.org/10.1016/j.atmosres.2016.10.017>.
- Thornton, J.A., Jaegle, L., McNeill, V.F., 2008. Assessing known pathways for HO<sub>2</sub> loss in aqueous atmospheric aerosols: regional and global impacts on tropospheric oxidants. *J. Geophys. Res.* 113 <https://doi.org/10.1029/2007JD009236>.
- Wang, P., Chen, Y., Hu, J., Zhang, H., Ying, Q., 2019. Attribution of tropospheric ozone to NO<sub>x</sub> and VOC emissions: considering ozone formation in the transition regime. *Environ. Sci. Technol.* 53, 1404–1412. <https://doi.org/10.1021/acs.est.8b05981>.
- Wang, S., Zhao, M., Xing, J., Ye, W.U., Zhou, Y.U., Lei, Y.U., Kebin, H.E., Lixin, F.U., Hao, J., 2010. Quantifying the air pollutants emission reduction during the 2008 olympic games in Beijing, 44. *Environmental Science & Technology*, pp. 2490–2496. <https://doi.org/10.1021/es9028167>.
- Wang, Y., Hao, J., Mcelroy, M.B., Munger, J.W., Ma, H., Chen, D., Nielsen, C.P., 2009. Ozone air quality during the 2008 Beijing Olympics: effectiveness of emission restrictions. *Atmos. Chem. Phys.* 9, 5237–5251.

- 5237–2009.
- Wei, W., Cheng, S., Li, G., Wang, G., Wang, H., 2014. Characteristics of ozone and ozone precursors (VOCs and NO<sub>x</sub>) around a petroleum refinery in Beijing, China. *J. Environ. Sci. China* 26, 332–342. [https://doi.org/10.1016/S1001-0742\(13\)60412-X](https://doi.org/10.1016/S1001-0742(13)60412-X).
- Wesely, M.L., 1989. Parameterization of surface resistances to gaseous dry deposition in regional-scale numerical models. *Atmos. Environ.* 23, 1293–1304. <https://doi.org/10.1016/j.atmosenv.2007.10.058>.
- Xu, R., Tang, G., Wang, Y., Tie, X., 2016. Analysis of a long-term measurement of air pollutants (2007–2011) in north China plain (NCP); impact of emission reduction during the Beijing Olympic games. *Chemosphere* 159, 647–658. <https://doi.org/10.1016/j.chemosphere.2016.06.025>.
- Yang, Y., Liao, H., Li, J., 2014. Impacts of the East Asian summer monsoon on interannual variations of summertime surface-layer ozone concentrations over China. *Atmos. Chem. Phys.* 14, 6867–6879. <https://doi.org/10.5194/acp-14-6867-2014>.
- Yang, Y., Liao, H., Lou, S., 2016. Increase in winter haze over eastern China in recent decades: roles of variations in meteorological parameters and anthropogenic emissions. *J. Geophys. Res.* 121 <https://doi.org/10.1002/2016JD025136>.
- Yu, H., Dai, W., Ren, L., Liu, D., Yan, X., Xiao, H., He, J., Xu, H., 2018. The effect of emission control on the submicron particulate matter size distribution in Hangzhou during the 2016 G20 summit. *Aerosol Air Qual. Res.* 18, 2038–2046. <https://doi.org/10.4209/aaqr.2018.01.0014>.
- Zhang, G., Xu, H., Wang, H., Xue, L., He, J., Xu, W., Qi, B., Du, R., Liu, C., Li, Z., 2020. Exploring the inconsistent variations in atmospheric primary and secondary pollutants during the G20 2016 Summit in Hangzhou, China: implications from observation and model. *Atmos. Chem. Phys.* 1–25. <https://doi.org/10.5194/acp-2019-1061>.
- Zhang, L., Liu, L., Zhao, Y., Gong, S., Zhang, X., Henze, D.K., Capps, S.L., Fu, T., Zhang, Q., Wang, Y., 2015. Source attribution of particulate matter pollution over North China with the adjoint method. *Environ. Res. Lett.* 10, 084011. <https://doi.org/10.1088/1748-9326/10/8/084011>.
- Zhang, L., Shao, J., Lu, X., Zhao, Y., Hu, Y., Henze, D.K., Liao, H., Gong, S., Zhang, Q., 2016. Sources and processes affecting fine particulate matter pollution over North China: an adjoint analysis of the Beijing APEC period. *Environ. Sci. Technol.* 50, 8731–8740. <https://doi.org/10.1021/acs.est.6b03010>.
- Zhu, J., Chen, L., Liao, H., Dang, R., 2019. Correlations between PM<sub>2.5</sub> and ozone over China and associated underlying reasons. *Atmosphere* 10, 352. <https://doi.org/10.3390/atmos10070352>.
- Zhu, J., Liao, H., 2016. Future ozone air quality and radiative forcing over China owing to future changes in emissions under the Representative Concentration Pathways (RCPs). *J. Geophys. Res.* 121, 1978–2001. <https://doi.org/10.1002/2015JD023926>.
- Zou, Y., Deng, X.J., Zhu, D., Gong, D.C., Wang, H., Li, F., Tan, H.B., Deng, T., Mai, B.R., Liu, X.T., Wang, B.G., 2015. Characteristics of 1 year of observational data of VOCs, NO<sub>x</sub> and O<sub>3</sub> at a suburban site in Guangzhou, China. *Atmos. Chem. Phys.* 15, 6625–6636. <https://doi.org/10.5194/acp-15-6625-2015>.